Effect of polluted water on soil and plant contamination by heavy metals in El-Mahla El-Kobra, Egypt

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Abstract. The discharge of untreated waste water in Zefta drain and drain no. 5 is becoming a problem for many farmers in the El-Mahla El-Kobra area, Egypt. The discharged water contains high levels of contaminants considered hazardous to the ecosystem. Some plants, soil, water, and sediment samples were collected from the El-Mahla El-Kobra area to evaluate the contamination by heavy metals. The results showed that the heavy metals, pH, sodium adsorption ratio (SAR), biochemical oxygen demand (BOD), and chemical oxygen demand (COD) in the water of Zefta drain and drain no. 5 exceeded permissible limits for irrigation. In rice and maize shoots grown in soils irrigated by contaminated water from Zefta drain and drain no. 5, the bioaccumulation factors for Cd, Pb, Zn, Cu, and Mn were higher than 1.0. The heavy metals content of irrigated soils from Zefta drain and drain no. 5 exceeded the upper limit of background heavy metals. In this study, the mean contaminant factor values of the drain no. 5 sediments revealed that Zn, Mn, Cu, Cd, Pb, and Ni > 6, indicating very high contamination. The bioaccumulation coefficient values of Cynodon dactylon, Phragmites australis, and Typha domingensis aquatic plants growing in Zefta drain are high. These species can be considered as hyperaccumulators for the decontamination of contaminated water.

1 Introduction

Contamination of soils by heavy metals, such as Cd, Ni, Zn, Pb, and Cu, has increased dramatically during the last few decades (Chibuike and Obiora, 2014) due to mining, smelting, manufacturing, use of agricultural fertilizers and pesticides, municipal waste, traffic emissions, and industrial effluents (Morgan, 2013; Chibuike and Obiora, 2014). Contamination of soils by heavy metals is now widespread (Al-Nagger et al., 2013). Land degradation caused by heavy metals has significant adverse effects on the environment and ecosystem worldwide (Li et al., 2013; Chen et al., 2015). Dispersion of heavy metals in irrigated soils and the plants that are growing results in the contamination of food that may be hazardous to humans and animals (Jolly et al., 2013). Heavy metals in effluents are poorly soluble in water, and cannot be degraded; they tend to accumulate in soils and subsequently accumulate in plants (Ghoneim et al., 2014). In addition, heavy metals persist in soil which then leach down into the groundwater and may induce enhanced antioxidant enzymatic activities in plants or become adsorbed with solid soil particles (Iannelli et al., 2002). According to Roy and McDonald (2013), carrots grown in soils contaminated by Cd have the potential to cause toxicological problems in men, women, and young children. Cd uptake by carrot roots was about 5 times more than the regulatory limits for men, 8 times more for women, and 12 times more for children. High levels of Cd in soil were identified as causing itai-itai disease in Toyama Prefecture, Japan; however, soil solution levels

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similarly high in Cd do not seem to cause health problems for people living in Shipham, England (Morgan, 2013). For the Cu-contaminated soils planted with tomato (*Solanum lycopersicum* L.), these values would range between 32.9 and 1696.5 mg kg⁻¹, depending on soil properties (Sacristán et al., 2015). Accumulation of toxic heavy metals in living plant cells results in various deficiencies, reduction of cell activities, and inhibition of plant growth (Farooqi et al., 2009).

Transfer of heavy metals from soils to plants is one of the key pathways for exposure of humans via the food chain. In order to assess risks to health associated with metals in soils, it is necessary to predict transfer of heavy metals from soil to tissues of plants for subsequent use in phytoremediation (Roy and McDonald, 2013; Ye et al., 2014). Heavy metal pollution is persistent, covert, and irreversible (Wang et al., 2011). This kind of pollution not only degrades the quality of the food crops, atmosphere, and water, but also threatens the health of human and animals (Dong et al., 2011). Excessive intake of the Pb to human body can damage the nervous, skeletal, endocrine, enzymatic, circulatory, and immune system (Zhang et al., 2012). The chronic effects of Cd consist of lung cancer, pulmonary adenocarcinomas, prostatic proliferative lesions, kidney dysfunction, bone fractures, and hypertension (Brevik et al., 2015).

Soil is a key part of the Earth system as it control the hydrological, erosional, biological, and geochemical cycles. The soil system also offers goods, services, and resources to humankind (Berendse et al., 2015; Brevik et al., 2015; Decock et al., 2015; Smith et al., 2015). This is why it is necessary to research how soils are affected by societies. Pollution is one of these damaging human activities, and we need more information and assessment of soil pollution (Mahmoud and El-Kader, 2015; Riding et al., 2015; Roy and Mcdonald, 2015; Wang et al., 2015). Soil degradation is now considered a challenge of a global dimension and is included in environmental policy frameworks. A prime example is the United Nations Convention to Combat Desertification (UNCCD), which recognizes the important role of soils in sustainable development and has anticipated the ambitious aim to achieve zero net land degradation by 2030 (UNCCD, 2012). Soils have been used to detect the deposition, accumulation, and distribution of heavy metals in different locations (Alirzayeva et al., 2006; Onder et al., 2007), but little quantitative information is available on the contamination of agri-
cultural soils in El-Malah El-Kobra, Egypt, by heavy metals. To close this knowledge gap, this study investigated the contamination of agricultural soil and plants by heavy metals in residential and industrial areas in El-Malah El-Kobra, Egypt.

2 Materials and methods

2.1 Site description, samples, and analysis

El-Malah El-Kobra (Fig. 1) is located at 30°34’N, 30°45’E. The dominant sources of heavy metal pollution are waste water irrigation, manure, and sediment applications for metallic ores. The El-Malah El-Kobra area is density populated (4.5 million 462 683 km⁻²) and contains 183 industrial factories such as for textiles, food, oil, and other industries. The quantity of industrial and municipal waste water is around 243 500 m³ day⁻¹ (107 500 m³ day⁻¹ municipal sewage and 136 000 m³ day⁻¹ industrial waste water), which discharges into Zefta drain (flow, 354 240 m³ day⁻¹) and drain no. 5 (flow, 265 248 m³ day⁻¹) without treatment, except 63 627 m³ day⁻¹ municipal waste water which can be treated at Dawakhlia plant.

Seventy representative soil samples (0–30 cm) in summer 2012 were collected from cultivated lands of El-Malah El-Kobra, Egypt, which are irrigated with drainage water from Zefta drain and drain no. 5, and 15 samples of soil were collected which is irrigated with water from Baher El Mlah canal (fresh water). The soil samples were air-dried and ground to pass through a 2 mm screen for chemical analysis. The soils’ pH was determined in a saturated soil paste extract (Richards, 1954). Calcium and magnesium levels were determined titrimetrically using versenate (Jackson, 1973). The level of sodium was determined using a flame photometer (Richards, 1954). The level of total carbonate was determined using the calcimeter as a CaCO₃ percentage according to Loeppert and Suarez (1996). The total heavy metals (Cd, Pb, Zn, Fe, Mn, Cu, and Ni) were measured by the atomic absorption spectrophotometer after the soil samples had digested concentrated mixtures of HNO₃ and HClO₄ acids (Page, 1982). Samples of rice and maize cultivated crops (age 65 days in summer 2012) that are grown in the studied soils, and three other aquatic plant species (Cynodon dactylon, Phragmites australis, and Typha domingensis) which are grown in Zefta drain, were also collected at different times. The plant samples were dried in an oven at 75°C for 72 h. The total heavy metals content in plant shoots was measured using the atomic absorption spectrophotometer after the plant samples had digested concentrated H₂SO₄ and H₂O₂ (Chapman and Pratt, 1961).

2.2 Transfer of heavy metals

The bioconcentration factor (BF) of each metal in plants was calculated by dividing the total content in the plants by the total content in soil (Brooks, 1998). In addition, 17 water samples were collected from Zefta drain and drain no. 5 at different times (March 2012 to March 2013) at about 20 cm below the water surface and were chemically analysed for pH, electrical conductivity (EC), sodium adsorption ratio (SAR), biological oxygen demand (BOD), chemical oxygen demand (COD), and heavy metals content (APHA, 2005). The bioaccumulation coefficients of each heavy metal in aquatic plants were calculated by dividing the total heavy metal content in aquatic plants by the concentration in water.

The contaminant factor (Cf) for soil is the ratio obtained by dividing the concentration of each heavy metal in the sediment by the background values (Håkanson, 1980).

$$\text{Cf} = \frac{C_{\text{Heavy metal}}}{C_{\text{Background}}}$$  \hspace{1cm} (1)

According to Håkanson (1980), the values of Cf < 1 indicate low contamination; 1 < Cf < 3 indicates moderate contamination; 3 < Cf < 6 indicates considerable contamination; and Cf > 6 indicates very high contamination.

3 Results and discussion

3.1 Effect of contaminated water on plant and soil contamination

Heavy metals content was higher in rice and maize shoots grown in the soil around the Zefta drain than the same crops in soil irrigated by water from drain no. 5 (Fig. 2). This was due to the high total heavy metal contents in these soils (Table 1). Maize shoots contained more Fe, Cd, Mn, and Pb than rice shoots, and this may be attributed to the planting of rice under flooded conditions. Under flooded conditions, Fe, Cd, Mn, and Pb could be precipitate as FeS₂, CdS, MnS, and PbS, respectively, due to the reducing conditions. Heavy metals content in rice and maize shoots exceeded the defined limits reported by Kabata-Pendias and Pendias (1992) and were above the levels acceptable for elemental composition of uncontaminated plant tissue. Alloway (1990) reported that in angiosperms, uncontaminated plant tissue contains 0.64,
Table 1. Total concentrations of heavy metals in soils irrigated by contaminated water from Zefta drain, drain no. 5, and Baher El Mlah canal.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Drain no. 5</th>
<th>Zefta drain</th>
<th>Baher El Mlah</th>
<th>Upper limit of total heavy metals in soils (Chen et al., 1992)</th>
</tr>
</thead>
<tbody>
<tr>
<td>pH</td>
<td>7.80–8.30</td>
<td>7.80–8.50</td>
<td>7.30</td>
<td>–</td>
</tr>
<tr>
<td>CaCO₃%</td>
<td>4.10–8.20</td>
<td>3.28–5.74</td>
<td>4.10</td>
<td>–</td>
</tr>
<tr>
<td>Fe mg kg⁻¹</td>
<td>1226–4989</td>
<td>1790–4757</td>
<td>933</td>
<td>–</td>
</tr>
<tr>
<td>Zn mg kg⁻¹</td>
<td>102–187</td>
<td>184–449</td>
<td>54</td>
<td>120</td>
</tr>
<tr>
<td>Mn mg kg⁻¹</td>
<td>341–800</td>
<td>172–853</td>
<td>264</td>
<td>–</td>
</tr>
<tr>
<td>Cu mg kg⁻¹</td>
<td>82–167</td>
<td>123–386</td>
<td>60</td>
<td>35</td>
</tr>
<tr>
<td>Cd mg kg⁻¹</td>
<td>13–28</td>
<td>21–33</td>
<td>11</td>
<td>3</td>
</tr>
<tr>
<td>Pb mg kg⁻¹</td>
<td>48–92</td>
<td>55–80</td>
<td>53</td>
<td>120</td>
</tr>
<tr>
<td>Ni mg kg⁻¹</td>
<td>55–133</td>
<td>104–164</td>
<td>31</td>
<td>60</td>
</tr>
</tbody>
</table>

2.4, 160, and 14 mg kg⁻¹ of Cd, Pb, Zn, and Cu, respectively. It is clear from Fig. 2 and Table 1 that the concentrations of Cd in rice and maize shoots are higher than other heavy metals compared with the maximum limits according to Kabata-Pendias and Pendias (1992). Li et al. (1994) found that plants absorb Cd more readily than other heavy metals and levels are often reached that are hazardous to human health before any stress symptoms appear. Chitdeshwari et al. (2002) reported that the use of sewage water increased the uptake of Cd and Cr in Amaranthus crops. Phosphate fertilizers were sources of Cd used in fertilization of rice and maize plants in this study area. Phosphate fertilizers were even measured to contain 200 mg Cd kg⁻¹ (Nziguheba and Smolders, 2008). The city of El-Mahla El-Kobra is densely populated and is the capital of the local textile industry. Large amounts of industrial and contaminated water are discharged directly into irrigation canals without treatment, which often contain heavy metals that contribute to metals’ enrichment in soil (Fakayode and Onianwa, 2002). In addition, rice and wheat ash fertilization is carried out in El-Mahla El-Kobra on a large scale (Abou-Sekkina et al., 2010). Application of ash to agricultural soils contributes significantly to the greater concentration of Cd in agricultural soil from El-Mahla El-Kobra. The higher concentration of Zn observed might be due to abrasion of tyres, barks, and Zn-containing compounds, which are used in some manufactured goods, such as paints, cosmetics, automobile tyres, and batteries (Imperato et al., 2003).

Table 2. Bioconcentration factors of heavy metals in maize and rice shoots grown in soils irrigated by waste water from Zefta drain, drain no. 5, and limits of heavy metals.

<table>
<thead>
<tr>
<th>Elements</th>
<th>Rice</th>
<th>Maize</th>
<th>Rice</th>
<th>Maize</th>
<th>Limits of heavy metals¹ mg kg⁻¹</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fe</td>
<td>0.29</td>
<td>0.35</td>
<td>0.54</td>
<td>0.34</td>
<td>–</td>
</tr>
<tr>
<td>Mn</td>
<td>3.40</td>
<td>2.97</td>
<td>1.71</td>
<td>1.51</td>
<td>300–500</td>
</tr>
<tr>
<td>Cu</td>
<td>0.59</td>
<td>1.54</td>
<td>1.75</td>
<td>1.83</td>
<td>20–100</td>
</tr>
<tr>
<td>Zn</td>
<td>2.82</td>
<td>1.71</td>
<td>1.44</td>
<td>1.32</td>
<td>100–400</td>
</tr>
<tr>
<td>Pb</td>
<td>6.73</td>
<td>6.83</td>
<td>5.26</td>
<td>5.66</td>
<td>30–300</td>
</tr>
<tr>
<td>Cd</td>
<td>6.14</td>
<td>6.45</td>
<td>6.55</td>
<td>2.26</td>
<td>5–30</td>
</tr>
</tbody>
</table>


The reason for an increasing soil pH may be attributed to high pH values in Zefta drain and drain no. 5 (Table 3). The soils irrigated by contaminated water from Zefta drain and drain no. 5 affect the total dissolved solids (TDSs) (Table 3). Indeed, in comparison to soils which are irrigated by water from Baher El Mlah, the TDS value is greater in Zefta drain and drain no. 5. These results were in agreement with Mollahoseini (2013) and Khaskhoussy et al. (2013), who reported that irrigating with sewage water increased soil salinity, exchangeable Na, K, Ca, Mg, and available P. In general, the concentrations of heavy metals in soils irrigated by water from Zefta drain and drain no. 5 exceeded the upper background limit of total heavy metals (Chen et al., 1992). Contents of Mn, Cd, and Ni in soils at Zefta drain were higher than in soils at drain no. 5, due to the high concentration of heavy metals in Zefta drain water (Table 3). The level of heavy metals of soils irrigated by water from Zefta drain and drain no. 5 was higher than the levels of the surrounding soils of Baher El Mlah canal. Similar results were reported by Chen et al. (1992), who found high levels of heavy metals in soils which are irrigated by polluted industrial waste water. These results coincided with El-Gendi et al. (1997) who indi-
cated that irrigating sandy soil in the Abou-Rawash area with drainage water increased total Cu, Zn, and Fe, which reached 125, 170, and 5 times that of the virgin soil in the same area. Both Cd and Pb levels in soils measured during this study were higher than those reported by Nassef et al. (2006) and Suciu et al. (2008). These differences might be related to different anthropogenic activities and concentrations of urbanization at each site.

3.2 Bioconcentration factors (BFs)

The BF values in the rice and maize shoots are presented in Table 2. In rice and maize grown in soils irrigated by water from Zefta drain and drain no. 5, the BF values for Cd, Pb, Zn, Cu, and Mn were higher than 1. This indicates that bioconcentrations of Cd, Pb, Cu, and Mn were high in the plants studied. Fe was an exception because its BF value was lower than 1, indicating low accumulation in studied plants. The BF values for Zn and Cu of rice shoots were higher than those reported by Nassef et al. (2006) and Suciu et al. (2008). These differences might be related to different anthropogenic activities and concentrations of urbanization at each site.

3.3 Quality of drainage water

Concentrations of BOD and COD ranged from 442 and 978 to 632 and 2445 mg L\(^{-1}\) in Zefta drain, while the BOD and COD value ranged from 540 and 882 to 723 and 2301 mg L\(^{-1}\) in drain no. 5 (Table 3). This water would be classified as high strength (Metcalf and Eddy, 2003). These results were in agreement with Pescond (1992) who reported that chemical properties of water include total dissolved solids (TDSs); BOD and COD showed higher values in untreated sewage water compared to groundwater. The BOD / COD ratio in Zefta drain and drain no. 5 ranged from 0.25 and 0.31 to 0.45 and 0.61, respectively. The waste water with a BOD / COD ratio below 0.50 contains some toxic components such as dyes and heavy metals (Linsley et al., 1992).

The average TDSs was 1016 in drain no. 5, 1130 in Zefta drain, and 334 mg L\(^{-1}\) in Baher El Mlah canal. The SARs in water of Zefta drain and drain no. 5 were above 12, which is considered a potential level for aggregate slaking, soil swelling, and clay dispersion, and thus reduction in hydraulic conductivity (Mace and Amrhein, 2001). The heavy metals in the two drains were higher than in the water of Baher El Mlah canal which could be attributed to discharge of industrial waste water into the two drains without treatment. The level of heavy metals exceeded the criteria limits for irrigation water (FAO, 2010; E.C.S, 48/1992). Similar results reported by Matloub and Mehana (1998) showed that sewage often has high values of temperature, pH, hardness, alkalinity, COD, TDS, NO\(_3\), NO\(_2\), Na, K, Ca, and Mg. Chitdeshwari et al. (2002) reported that increased levels of sewage water increased the uptake of Cd and Cr in Amaranthus crops.

3.4 Heavy metal concentrations in sediments

The high heavy metal concentrations in sediments of drain no. 5 (Table 4) can be attributed to higher pH in water which can form ions of insoluble precipitates. The measured concentrations of heavy metals are higher than US Environmental Protection Agency (EPA) toxicity values (US EPA, 1999). Similar findings were reported by Thuy et al. (2007) who found that heavy metals in sediments of five canals which received untreated industrial waste water exceeded the US EPA toxicity levels. The partitioning of heavy metals between sediment and water can be expressed by distribution coefficient (Kd) values (L kg\(^{-1}\)). Kd values of sediment samples were the highest for Zn, Cd, and Mn, and lowest for Pb, Cu, and Ni. The higher Kd value indicates that the sorption of heavy metals by sediments was strong (Salomons and Forstner, 1980). Sediments are both carriers and potential sources of contaminants in aquatic systems, and these materials also affect groundwater quality and agricultural products when disposed of on land. In this study, the mean Cf values of the drain no. 5 sediments revealed that Zn, Mn, Cu, Cd, Pb, and Ni > 6, indicating very high contamination due to the direct discharge of waste water from the residential and industrial areas.
Table 3. The chemical analysis of water collected from the Zefta drain, drain no. 5, and Baher El Mlah canal.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Units</th>
<th>Drain no. 5</th>
<th>Zefta drain</th>
<th>Baher El Mlah</th>
<th>Water criteria for irrigation water</th>
</tr>
</thead>
<tbody>
<tr>
<td>pH</td>
<td>–</td>
<td>9.80</td>
<td>12.2</td>
<td>7.20</td>
<td>6.5–8.4</td>
</tr>
<tr>
<td>TDS</td>
<td>mg L(^{-1})</td>
<td>1016</td>
<td>1130</td>
<td>334</td>
<td>2000</td>
</tr>
<tr>
<td>SAR</td>
<td>–</td>
<td>17.3</td>
<td>18.2</td>
<td>6.00</td>
<td>6–12</td>
</tr>
<tr>
<td>BOD</td>
<td>mg L(^{-1})</td>
<td>540–723</td>
<td>442–632</td>
<td>0.00</td>
<td>40(^1)</td>
</tr>
<tr>
<td>COD</td>
<td>mg L(^{-1})</td>
<td>882–2301</td>
<td>978–2445</td>
<td>0.00</td>
<td>60(^1)</td>
</tr>
<tr>
<td>Fe</td>
<td>mg L(^{-1})</td>
<td>0.09</td>
<td>0.56</td>
<td>0.01</td>
<td>5.0</td>
</tr>
<tr>
<td>Zn</td>
<td>mg L(^{-1})</td>
<td>0.02</td>
<td>0.037</td>
<td>0.00</td>
<td>2.00</td>
</tr>
<tr>
<td>Mn</td>
<td>mg L(^{-1})</td>
<td>0.68</td>
<td>2.91</td>
<td>0.03</td>
<td>0.20</td>
</tr>
<tr>
<td>Cu</td>
<td>mg L(^{-1})</td>
<td>0.15</td>
<td>0.28</td>
<td>0.12</td>
<td>0.20</td>
</tr>
<tr>
<td>Cd</td>
<td>mg L(^{-1})</td>
<td>0.03</td>
<td>0.07</td>
<td>0.001</td>
<td>0.01</td>
</tr>
<tr>
<td>Pb</td>
<td>mg L(^{-1})</td>
<td>1.05</td>
<td>0.18</td>
<td>0.05</td>
<td>5.00</td>
</tr>
<tr>
<td>Ni</td>
<td>mg L(^{-1})</td>
<td>0.12</td>
<td>0.31</td>
<td>0.02</td>
<td>0.20</td>
</tr>
</tbody>
</table>

\(^1\) Egyptian Chemical Standards (48/1992).

Table 4. Average heavy metal concentrations, contaminant factor, and distribution coefficients (Kd) in sediments of drain no. 5 compared with toxicological reference values (US EPA, 1999).

<table>
<thead>
<tr>
<th>Element</th>
<th>Conc. (mg kg(^{-1}))</th>
<th>mean ± SD</th>
<th>Et</th>
<th>Cf</th>
<th>Kd (L kg(^{-1}))</th>
</tr>
</thead>
<tbody>
<tr>
<td>Zn</td>
<td>647.5 ± 36.7</td>
<td>110</td>
<td>6.25</td>
<td></td>
<td>32375.0</td>
</tr>
<tr>
<td>Mn</td>
<td>2125.0 ± 74.3</td>
<td>–</td>
<td>12.7</td>
<td></td>
<td>3125.0</td>
</tr>
<tr>
<td>Cu</td>
<td>425.0 ± 12.4</td>
<td>16</td>
<td>4.25</td>
<td></td>
<td>2833.3</td>
</tr>
<tr>
<td>Cd</td>
<td>97.5 ± 4.6</td>
<td>0.6</td>
<td>9.55</td>
<td></td>
<td>3250.0</td>
</tr>
<tr>
<td>Pb</td>
<td>145.0 ± 4.5</td>
<td>31</td>
<td>4.80</td>
<td></td>
<td>138.10</td>
</tr>
<tr>
<td>Ni</td>
<td>195.0 ± 9.8</td>
<td>–</td>
<td>7.33</td>
<td></td>
<td>1625.0</td>
</tr>
</tbody>
</table>

Et: US EPA toxicity reference value; Cf: contaminant factor; Kd: distribution coefficients (L kg\(^{-1}\)).

3.5 Bioaccumulation coefficients of aquatic plants

The bioaccumulation coefficients of heavy metals in plants of *Cynodon dactylon*, *Phragmites australis*, and *Typha domingensis* grown in Zefta drain are shown in Fig. 3. The bioaccumulation coefficients of metals in *Cynodon dactylon* were higher than in *Phragmites australis* and *Typha domingensis*. These plant species can be considered as hyperaccumulators, and are used for the decontamination of contaminated water. The use of plants for decontamination of polluted waters has been described as rhizofiltration (Brooks, 1998). The three species would be useful for phytoremediation of contaminated water in a particular area. Bonanno (2013) showed that *Phragmites australis* and *Typha domingensis* species may be used as biomonitors of trace element contamination in sediment. Overall, *T. domingensis* and *P. australis* showed a greater capacity of bioaccumulation as well as a greater efficiency of element removal than *A. donax*. In particular, *T. domingensis* and *P. australis* may be used for Hg phytoextraction and phytostabilization; the former also acted as a hyperaccumulator for trace elements’ phytoextraction and phytostabilization. In contaminated wetlands, the presence of *T. domingensis* and *P. australis* may increase the general retention of trace elements. Wafaa et al. (2010) demonstrated that *Phragmites australis* and *Tamarix aphillya* species are significant as vegetation filters and for cleaning the soils from contamination by heavy metals by phytoextraction. Antioxidant thiol compounds were probably involved in the mechanisms used by *P. australis* to alleviate metal toxicity. As *P. australis* is considered suitable for phytostabilizing metal-contaminated sediments, understanding its tolerance mechanisms to toxic metals is important to optimize the conditions for applying this plant in phytoremediation (Rocha et al., 2014).

4 Conclusions

Delta drains often receive high amounts of organic and inorganic pollutants from industrial and domestic waste water as well as diffuse agricultural drainage. High priority should be given to Zefta drain and drain no. 5 sites which receive high loads of pollutants. This was confirmed by the lower water quality and soils polluted by heavy metals in the El-Mahla El-Kobra area. Industrial and municipal waste water sources in El-Mahla El-Kobra area must be treated before being discharged in Zefta drain and drain no. 5. Using agricultural soils contaminated by heavy metals in the cultivation of rice and maize crops for human consumption may result in health hazards.
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References


Egyptian Chemical Standards: Protection of the Nile River and water stream from pollution, Ministry of Irrigation, Cairo, Egypt, low No. 48/1992.


